Linkages between physicochemical status and hydromorphology in Greek lakes under WFD policy

E. Mavromati^{1*}, I. Kagalou², D. Kemitzoglou¹, A. Apostolakis¹ and V. Tsiaoussi¹

¹ The Goulandris Natural History Museum – Greek Biotope Wetland Centre (EKBY), 14th km Thessaloniki - Mihaniona, 57001 Thermi, Greece

² Department of Civil Engineering, School of Engineering, Democritus University of Thrace, University Campus, Kimmeria, 67100, Xanthi, Greece

*e-mail: emavromati@ekby.gr

Abstract: The national monitoring network of waters in Greece, in the context of Water Framework Directive, has been operational since 2012. It comprises 50 lake water bodies, both natural and artificial (24 and 26 respectively). The aims of the study are i) to present the main features of Greek lakes; ii) to position them according to hydromorphological and physicochemical data from the first period of monitoring and iii) to investigate how the hydromorphological features influence water quality. Greek lakes are discerned in relation to hydromorphological aspects such as mean depth, volume, surface area, catchment area. Impacted and unimpacted lakes are separated when positioned along physical and chemical attributes, such as Secchi depth, TP and ion concentrations. The dataset covers a gradient of eutrophication, which seems to be the main anthropogenic pressure on lakes in Greece and probably one of the most widespread anthropogenic pressures on lakes across Europe. Catchment areas of the Greek lakes have undergone substantial agricultural, industrial, and urban development over recent years, leading to eutrophication and hydromorphological alterations. Management measures at a catchment scale aiming to control land uses would be needed in order to maintain and/or improve water quality in Greek lakes, supplemented by site-specific measures when appropriate. The monitoring network provides baseline data that will allow the assessment of status and trends of Greek lakes.

Key words: WFD, Greek lakes, hydromorphology, water quality

1. INTRODUCTION

The Water Framework Directive (WFD) foresees the operation of monitoring programmes for surface waters and groundwater in order to assess their ecological and chemical status (article 8). Comparable pan-European datasets, with hydromorphological, physicochemical and biological information are used to assess ecological status of surface waters and improve our knowledge on the structure and function of these ecosystems (Hering et al. 2010; Moe et al. 2013).

The complex geomorphology of Greece, the uneven distribution of precipitation and different regional demands for irrigation and drinking water are some of the characteristics that affect the hydromorphological features of Greek lakes (Zacharias 2002); this in turn strongly governs various ecosystem functions (Schindler and Scheurell 2002). According to Nõges (2009), nutrient inputs are profoundly affected by catchment and lake area through surface and groundwater drainage while nutrient retention, a fundamental property of every aquatic ecosystem, is linked to its geomorphological, hydrological, edaphic and biotic characteristics (Hansson et al. 2005). Furthermore, intensive agriculture activities, domestic and industrial inflows have changed the surface water distribution and use (Skoulikidis et al. 1998).

In line with WFD, Greece had to establish a national monitoring network (JMD 140384/2011); EKBY now operates the monitoring network for lakes in Greece, establishing baseline data to assess their ecological status and trends. The aims of the study are i) to present the main features of Greek lakes, ii) to position them according to morphological and physicochemical data from the first period of monitoring and iii) to investigate how the hydromorphological features influence water quality.

2. MATERIALS AND METHODS

2.1 Study area and data sets

The Greek National Water Monitoring Network for lakes comprises 50 water bodies, both natural and reservoirs. The heterogeneity of Greek geology, climate and topography has resulted in a lake dataset with a broad range of morphological, physical and chemical characteristics. Fifty-three stations have been established, 27 of which are subject to surveillance monitoring and 26 to operational. Most water bodies have one sampling station each, except transboundary lakes Megali Prespa, Mikri Prespa and Doirani, which have two sampling stations each (Figure 1).

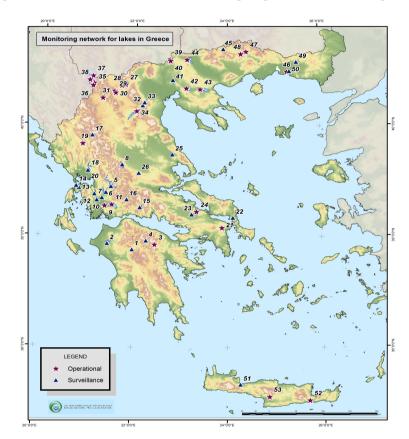


Figure 1. Monitoring network for lakes in Greece.

Physicochemical features were measured seasonally and each month during growing season (May to October). Samples were taken from euphotic zone [2,5xSecchi Depth (SD)]. In particular: Transparency was measured with a Secchi disc. Temperature, pH, oxygen, conductivity and Total Dissolved Solids were measured with portable meters from ThermoScientific instruments. Filtrations for Chlorophyll a measurements were carried out through Whatman GF/F glass-fiber filters and Chlorophyll a was estimated using standard methods (APHA 2012). Analyses for Na⁺, K^+ , Cl⁻ and SO4⁻ were carried out with the use of Ion Chromatography (ISO 14911 1999, ISO 10304-01 2009). Analyses of Total Phosphorus (TP) were carried out in unfiltered water using ascorbic method following persulfate digestion (APHA 2012). Data used for this study were taken from samplings carried out in 2015. To estimate the catchment area impact on water bodies, the Schindler's ratio was calculated as the ratio between the catchment and the surface area to lake volume (Stefanidis and Papastergiadou 2012). Intensive agriculture was measured as the sum of the Corine Land Cover categories corresponding to a high potential impact from agricultural activities, such as arable land (including irrigated land), permanent crops (with associated annual crops), vineyards, orchards, olive groves and complex cultivation patterns. Morphological parameters of transboundary lakes were estimated including data from sharing countries.

2.2 Statistical Analysis

All statistical analysis, including Boxplots, Pearson's Correlations and Principal Component Analysis (PCA), was conducted using R Statistical Software (R Core Team 2014).

Data analysis was performed only in stations sampled for at least two months over growing season. Very shallow natural lakes Koroneia, Saltini, Pikrolimni, Stymfalia and Dystos were excluded from analysis, due to either outlier values or lack of data caused by drought. Values below the quantification limit were replaced with the value of quantification. Variables were tested for skewness and log transformed when needed. Correlation Analysis was conducted in order to elucidate possible relationships between parameters. PCA was used to interpret the major patterns of variation within the data and ordinate the lakes with respect to selected hydromorphological and physicochemical variables.

3. RESULTS

Correlation analysis between TP and Schindler's Ratio revealed a significant relationship (p<0.05), as the two variables were positively correlated in both reservoirs and natural lakes (r=0.45 and r=0.41, respectively) (Table 1). The relationship between TP and Mean depth was statistically significant in both cases (p<0.01) and the r values confirmed the strong negative correlation between the two variables.

	Natural Lakes (n=20) LogTP	Reservoirs (n=23) LogTP
LogSchindler's Ratio	0.41*	0.45*
LogMeanDepth	-0.60**	-0.79**
* significant at $p < 0.05$, ** significant at $p < 0.01$		

Table 1. Pearson correlation between TP, Schindler's Ratio and Mean Depth

Boxplots in Figure 2 show TP and Chlorophyll a data (growing season mean) plotted against Intensive Agriculture and Population Density groups. Intensive agriculture activities affirmed their high impact on TP loading in water bodies, both reservoirs and natural lakes. Population density

showed the same positive relationship with TP, as expected. PCA revealed the ordination of Greek natural lakes in two PCs, with eigenvalues above one $(\lambda_1=3.11 \text{ and } \lambda_2=1.42, \text{ respectively})$, which captured 91.51% of total variance (Figure 3). All variables had high scores on the major PCA axis I, with lake volume standing out and negatively correlated with the axis. The second axis was positively associated with Schindler's ratio and catchment area. In reservoirs, the two first axes captured 81.94% of total variance, with lake volume and catchment area, being strongly correlated with PC1 whereas PC2 was only affected by Schindler's ratio.

With regard to the physicochemical aspects, the two principal components of the PCA explained 84.04% of the data variance, in natural lakes (Figure 4). The first axis (λ_1 =2.13) was driven mainly by electrical conductivity and sodium and less by calcium ions. The second axis was associated with Secchi depth and TP, revealing a clear gradient from deep, clear water bodies to shallow more impacted ones. The same gradient was evident in reservoirs, as TP and Secchi depth were the principal drivers of the second axis, with the former being now the strongest one. Electrical conductivity, sulphate and sodium ions were the most influential parameters of axis 1, separating the two reservoirs from Crete, T.L. Faneromenis and T.L. Bramianon. One reservoir, T.L Pournariou is separated from the others, as it is highly associated with calcium ions.

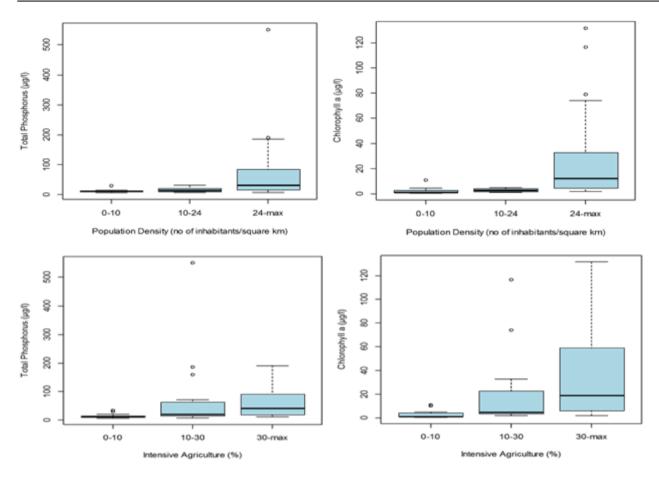


Figure 2. Boxplots of Intensive Agriculture and Population Density groups in relation to TP and Chorophyll a.

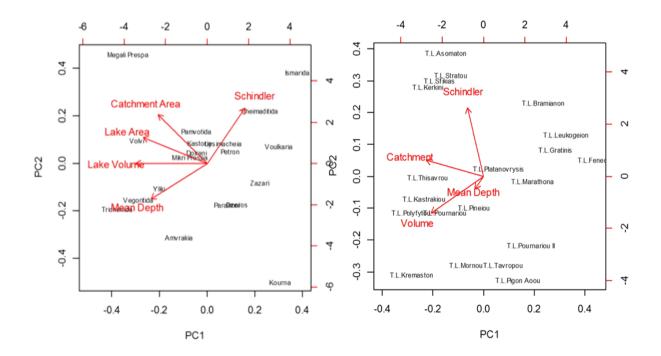


Figure 3. PCA biplots of selected morphological variables in Greek lakes dataset.

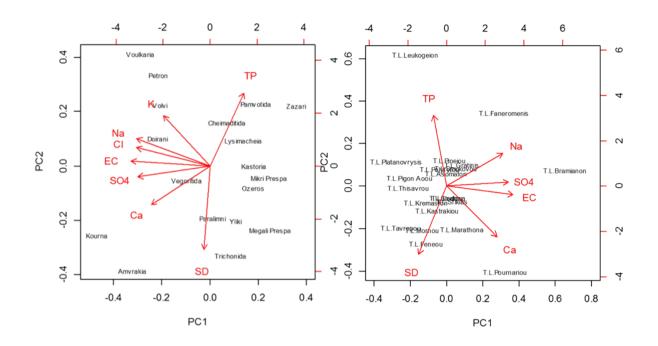


Figure 4. PCA biplots of selected physicochemical variables in Greek lakes dataset.

4. DISCUSSION

Totally non-impacted lakes are scarce in Europe (Poikane et al. 2010); the main pressures include nutrient inputs from point and non-point sources, water abstraction and morphological changes (EEA 2012). The majority of Greek lakes have been modified to some degree by the same pressures, as described in the River Basin Management Plans (http://wfd.ypeka.gr/).

The strong relationship between Schindler's ratio and mean depth with TP reflects the influence of catchment area's size and lake morphological parameters to lake water quality and further to eutrophication. Lake morphometry parameters (e.g. mean depth) indicate the buffer capacity to nutrient inputs (Nõges 2009). Furthermore, as Kolada et al. (2005) suggested, lakes with large catchment area are more susceptible to degradation. In such cases, we suggest that efficient restoration and mitigation strategies could focus rather at catchment level than in lake itself. Land uses and population density in the catchment, further influence water quality, as boxplots revealed a pattern of high population densities and intensive agriculture related to high TP and chlorophyll a values. The impacts of land use changes are complex. They may lead to a gradual conversion (e.g. habitat alterations, invasion of alien species) or total modifications (e.g. turbid conditions, food web changes) (Moss 1998; Zogaris et al. 2009). Similarly, the catchment areas of the Greek lakes have undergone substantial agricultural, industrial, and urban development over recent years, leading to eutrophication and hydromorphological alterations.

With regard to hydromorphology of Greek natural lakes, lake volume and mean depth stand out and highlight three large deep lakes, Yliki, Vegoritida, Trichonida. Schindler's ratio and catchment area reveal a gradient of increasing catchment area impact along the second axis of PCA. Lakes Trichonida, Vegoritida and Amvrakia have by far the lowest Schindler's ratio, whereas Ismarida has the highest value. Urban lakes Kastoria and Pamvotida along with Cheimaditida, Lysimacheia and Petron are clustered on the top right of the plot and distinguished by high values of Schindler's ratio. Using the reservoirs' dataset, the mountainous ones with low Schindler ratio impact, such as Aoou and Tavropou, are discriminated from lowland reservoirs with high catchment area impact. Furthermore, lake volume grouped large lakes Thisavrou, Kastrakiou, Kremaston, Polyfytou and Pournariou.

The physicochemical variables in PCA plots revealed a similar pattern in both reservoirs and

natural lakes. A gradient of increasing TP along the second component, separate clear reservoirs and deep natural lakes with high Secchi depth from shallow, impacted lakes with low Secchi depth. This describes the sensitivity of shallow lakes to resuspension process and further to nutrient loading. Urban lake Pamvotida, along with lakes Zazari, Cheimaditida and Lysimacheia are clustered on the top left quadrat of the plot, highly associated with TP. Nutrient loading is the result of household wastewater inputs and intense agriculture activities (Skoulikidis et al. 1998). Specifically, regarding the urban lakes, as lakes Pamvotida and Kastoria, it is well documented that they tend to receive high nutrient loads and show higher trophic status than non-urban ones (Naseli-Flores 2008); thus site-specific mitigation measures should be implemented. Water bodies, Kourna and Amvrakia, are positioned together on the bottom left of the plot, as they are highly associated with calcium, with the latter also related with sulphates. In Amvrakia, a hard water lake, several studies have been reported high concentrations of these ions, mainly due the presence of gypsum of the surroundings (Danielidis et al. 1995; Skoulikidis et al. 1998).

Findings suggest that intervention at a catchment scale aiming to control land uses would be needed in order to maintain and/or improve water quality in Greek lakes, supplemented by site-specific mitigation measures when appropriate. The sustainable and cost-effective management of Greek lakes requires a proper understanding of the relationships between morphometry and physicochemical variables. Continuous surveys would be useful in providing long-series data to be able to link hydromorphology and physicochemical conditions to ecological processes and biological responses.

ACKNOWLEDGEMENTS

This study has been prepared in the framework of the National Water Monitoring Network for lakes, according to the Joint Ministerial Decision 140384/2011. The acquisition of data was financed by ERDF - Operational Program Environment and Sustainable Development and national funds.

REFERENCES

- APHA, 2012. Standard methods for the examination of water and wastewater. American Public Health Association, American Water Works Association, Water Environment Federation, 22nd Ed.
- Danielidis D.B., M. Spartinou and A. Economou-Amilli, 1995. Limnological survey of Lake Amvrakia, western Greece. Hydrobiologia, 318: 207-218.
- European Commission, 2000. Directive 2000/60/EC of the European Parliament and of the Council of 23 October 2000 establishing a framework for Communities in the field of water policy, Official Journal of the European Communities, L 327/1, 22.12.2000.
- European Environment Agency, 2012. European Waters Assessment of Status and Pressures. European Environment Agency, Copenhagen.
- Hansson, L., C. Brönmark, N. P. Andersm and K. Åbjörnsson, 2005. Conflicting demands on wetland ecosystem services: nutrient retention, biodiversity or both? Freshwater Biology, 50(4): 705–714.
- Hering, D., A. Borja, J. Carstensen, L. Carvalho, M. Elliott, C. K. Feld, A.S. Heiskanen, R.K. Johnson, J. Moe, D. Pont, A.L. Solheim, W. van de Bund, 2010. The European Water Framework Directive at the age of 10: A critical review of the achievements with recommendations for the future. Science of the Total Environment, 408: 4007-4019.
- ISO 14911, 1999. Water quality Determination of dissolved Li⁺, Na⁺, K⁺, NH4⁺, Mg²⁺, Ca²⁺, Mn²⁺, Sr²⁺ and Ba²⁺ using ion chromatography Method for water and waste water. pp. 18
- ISO 10304-01, 2009. Water Quality Determination of dissolved anions by liquid chromatography of ions-Part 1: Determination of bromide, chroride, fluoride, nitrate, nitrite, phosphate and sulfate. pp. 15
- Joint Ministerial Decision No 140384/2011 (Official Gazette 2017 II 09.09.2011) establishing the "National Monitoring Network for the quality and the quantity of waters".
- Kagalou, I. and I. Leonardos, 2009. Typology, classification and management issues of Greek lakes: Implication of the water framework directive (2000/60/EC). Environmental Monitoring and Assessment, 150: 469-484.
- Kolada A., H. Soszka, D. Sydzik and M. Golub, 2005. Abiotic typology of Polish lakes. Limnologica, 35: 145-150.
- Ministry of Environment, Special Secretariat for Waters. Approved River Basin Management Plans (http://wfd.ypeka.gr/index.php?option=com_content&task=view&id=113&Itemid=19)
- Moe, S. J., A. Schmidt-Kloiber, B. J. Dudley and D. Hering, 2013. The WISER way of organising ecological data from European rivers, lakes, transitional and coastal waters. Hydrobiologia, 704: 11-28.

Moss B., 1998. Ecology of freshwaters: Man and Medium, Past to Future. 3rd Ed., Blackwell Publishing, pp. 14–35.

- Nõges, 2009. Relationships between morphometry, geographic location and water quality parameters of European lakes. Hydrobiologia, 633: 33-43.
- Naselli-Flores L., 2008. Urban Lakes: Ecosystem at risk, worthy of the best care. Proceedings Taal2007: The 12th World Lake Conference, pp. 1333-1337.
- Poikane S., M.H. Alves, C. Argillier, M. van den Berg, F. Buzzi, E. Hoehn, C. de Hoyos, I. Karottki, C. Laplace-Treyture, A.L. Solheim, J. Ortiz-Casas, I. Ott, G. Phillips, A. Pilke, J. Pádua, S. Remec-Rekar, U. Riedmüller, J. Schaumburg, M. L. Serrano, H. Soszka, D. Tierney, G. Urbanič, and G. Wolfram. 2010. Defining Chlorophyll a reference conditions in European lakes. Environmental Management, 45(6): 1286–1298.
- R Core Team, 2014. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria.

Schindler D.E. and M.D. Scheuerell, 2002. Habitat coupling in lake ecosystems. Oikos 98: 177-189.

- Skoulikidis, N., Bertachas I. and T. Koussouris, 1998. The environmental state of fresh water resources in Greece (rivers and lakes). Environmental Geology, 36: 1–17.
- Stephanidis, K. and E. Papastergiadou, 2012. Relationships between lake morphology, water quality and aquatic macrophytes in lakes of Western Greece. Fresenius Environmental Bulletin 21: 3018-3026.
- Zacharias, I., I. Bertachas, N. Skoulikidis and T. Koussouris, 2002. Greek Lakes: Limnological overview. Lakes & Reservoirs: Research and Management, 7: 55-62.
- Zogaris S., A. Economou and P. Dimopoulos, 2009. Ecoregions in the Southern Balkans: should their boundaries be revised? Environmental Management, 43: 682–697.